

# Tracing the origin of reservoir sediments using magnetic properties in Southeastern Brazil

## Traçando a origem de sedimentos de reservatórios usando propriedades magnéticas no Sudeste do Brasil

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### Highlights:

A magnetic tracing analysis to link upstream areas to reservoir sedimentation is used in a subtropical environment. Tropical weathering conditions promote magnetic variability between soils. Magnetic parameters can identify possible sediment sources in a tropical environment in order to reduce water erosion impacts.

### Abstract

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Determining the origin of eroded soil is essential to design effective soil erosion control strategies which preserve the soil resource, enhance agricultural productivity, and reduce the negative impacts of soil erosion, in-field and off-field. Magnetic properties have been widely used in temperate environments to identify sediment sources, pathways and links, but there have been very few applications in tropical and subtropical environments. Therefore, in this paper we investigated reservoir sediment sources in the Upper Grande River Basin, Southeastern Brazil, using sediment tracing techniques based on magnetic parameters (low and high frequency magnetic susceptibility, frequency dependent susceptibility). The different parent materials and subtropical weathering conditions resulted in soils having different Fe oxide minerals and Fe oxide contents, promoting magnetic variability that allowed comparison and identification of possible sources of reservoir sediments in order to reduce water erosion impacts. The results indicate the suitability of magnetic properties as a tracer for soil erosion studies in tropical environments.

**Key words:** Natural resources. Sediment sources. Soil erosion. Tropical environment.

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## Resumo

Determinar a origem de solos erodidos é essencial para projetar estratégias efetivas de controle da erosão do solo que preservem os recursos do solo, aumentem a produtividade agrícola e reduzam os impactos negativos da erosão do solo, em campo e fora dele. As propriedades magnéticas têm sido amplamente utilizadas em ambientes temperados para identificar fontes, vias e elos de sedimentos, mas existem pouquíssimas aplicações em ambientes tropicais. Portanto, neste trabalho investigamos fontes de sedimentos de reservatórios na Bacia do Alto Rio Grande, Sudeste do Brasil, utilizando técnicas de rastreamento de sedimentos baseadas em parâmetros magnéticos (susceptibilidade magnética de baixa e alta frequência, susceptibilidade dependente da frequência). Os diferentes materiais de origem e condições de intemperismo tropical resultaram em solos com diferentes minerais de óxido de Fe e teores de óxidos de Fe, promovendo variabilidade magnética que permitiu a comparação e identificação de possíveis fontes de sedimentos de reservatórios para reduzir os impactos da erosão hídrica. Os resultados indicam a adequação das propriedades magnéticas como um traçador para estudos de erosão do solo em ambientes tropicais.

**Palavras-chave:** Ambiente tropical. Erosão do solo. Fontes de sedimentos. Recursos naturais.

### Introduction

Soil erosion by water is a major consequence of land degradation and results in a consequent reduction in agricultural productivity worldwide (Pimentel et al., 1995). Negative effects of water-driven soil erosion include off-field impacts such as changes in water quality, disruption to biological processes, and siltation of streams and reservoirs (Jain & Singh, 2003; Batista et al., 2017; Bostanmaneshrad et al., 2018).

In Brazil, water erosion has increased exponentially due to agricultural expansion (Oliveira, Nearing, & Wendland, 2015; Anache, Wendland, Oliveira, Flanagan, & Nearing, 2017). Hence, consequent sedimentation due to accelerated erosion of fine-grained sediment leads to a reduction in water availability and a deterioration in water quality (Araujo, Güntner, & Bronstert, 2006; Batista et al., 2017). Furthermore, finding the origin of the eroded soil and sediments can contribute to the preservation of natural resources and mitigate the off-field impacts by enabling targeted soil erosion and sediment control strategies (Collins, Walling, Sickingabula, & Leeks, 2001).

Research into soil and sediment losses by erosion have traditionally used standard erosion plots for

monitoring water-driven erosion (Wischmeier & Smith, 1978; Anache et al., 2017). Certainly, erosion rate quantification using these methods is an essential part of monitoring agricultural practices in order to determine how soil management systems affect water and sediment runoff (Zhang, Nearing, Garbrecht, & Steiner, 2004; Bispo et al., 2017; Le Gall et al., 2017; Saran, Meneghine, Célico, Pinheiro, & Alves, 2017). However, although capable of providing useful information, such plots have limitations in terms of data representativeness, spatial and temporal resolution, and cost (Armstrong, Quinton, & Maher, 2012; Guzmán, Quinton, Nearing, Mabit, & Gómez, 2013; Deasy, Titman, & Quinton, 2014; Batista, Davies, Silva & Quinton, 2019).

Alternatively, soil tracers can be used to identify eroding areas (Guzmán et al., 2013; Collins et al., 2017). Tracer methodologies are used to determine soil loss ratio or sediment production and to track soil redistribution through the landscape (Guzmán et al., 2013). One of the main groups of tracing studies is based on the principle that suspended sediments retain some of the properties acquired at their origin, such that a sediment sample transported through the landscape can be compared to potential sources within the watershed. This technique is

called “fingerprinting” (Collins & Walling, 2002; Armstrong et al., 2012; Walling, 2013; Collins et al., 2017).

Various sediment properties have been used in sediment tracing and finger-printing studies, including chemical (organic C, inorganic C, total C, C/N, pH, extractable Ca, extractable Mg, extractable K, extractable Na,  $\delta^{13}\text{C}$ ,  $\delta^{15}\text{N}$ , total Si, Al, Ag, Bi, Cd, Cr, Hg, Fe, Ca, Mg, Mn, Na, K, Ti, P, Zn, Sr, Pb, Ni, Cu, As, Mo, Sn, U, Pb, Sb, Sn, inorganic P, organic P, total P,  $^{204}\text{Pb}$ ,  $^{206}\text{Pb}$ ,  $^{207}\text{Pb}$ ,  $^{208}\text{Pb}$ ), physical (sand, clay and silt fraction, water dispersible clay, aggregation index,  $^{137}\text{Cs}$ ,  $^{40}\text{K}$ ,  $^{226}\text{Ra}$ , unsupported  $^{210}\text{Pb}$ ), biological (sterol rations, *E. coli*, Enterococci bacterial signatures) and magnetic properties (Guzmán et al., 2013).

As a low cost alternative, magnetic properties are widely used in temperate environments to identify sediment sources, pathways and links. In fact, different natural materials display different magnetic properties, often enabling magnetic mineral identification, helping soil classification and identification of soil-forming processes, as well as attribution of the eroded sediment source (Walden, Oldfield, & Smith, 1999). Several factors can directly influence the magnetic variables that can be used for tracing sediments: soil parent material influences the primary, detrital mineralogy type and content of iron oxides: relief, landscape position, vegetation, weathering conditions and climate can all subsequently influence the pedogenic magnetic characteristics of the soil (Maher, 1998; Blundell, Dearing, Boyle, & Hannam, 2009). Consequently, soils, sediments and rocks from different locations in a catchment are characterized by different magnetic properties that can be relatively easily quantified (Maher, Watkins, Brunskill, Alexander, & Fielding, 2009).

Soil magnetic properties as sediment fingerprints have been widely applied in temperate zones (Royall, 2003; Hatfield & Maher, 2009; Armstrong et al., 2012). However, there have been very few

applications of magnetic sediment tracing in the tropics. In Brazil, magnetic sediment fingerprinting has been successfully applied to investigate environmental processes, indicating the potential for the technique to improve soil and sediment erosion control. In addition, magnetic mineralogy fingerprinting was applied in Southern Brazil and detected a shift in sediment delivery at the estuary of the Paraná River from distinct sources of sediments, from fine-grained magnetite to coarse-grained hematite derived from basalt (Mathias, Nagai, Trindade, & Mahiques, 2014). Moreover, magnetic susceptibility was also successfully used as a predictor of erodibility factors in the modeling process for large tropical areas (Barbosa et al., 2019).

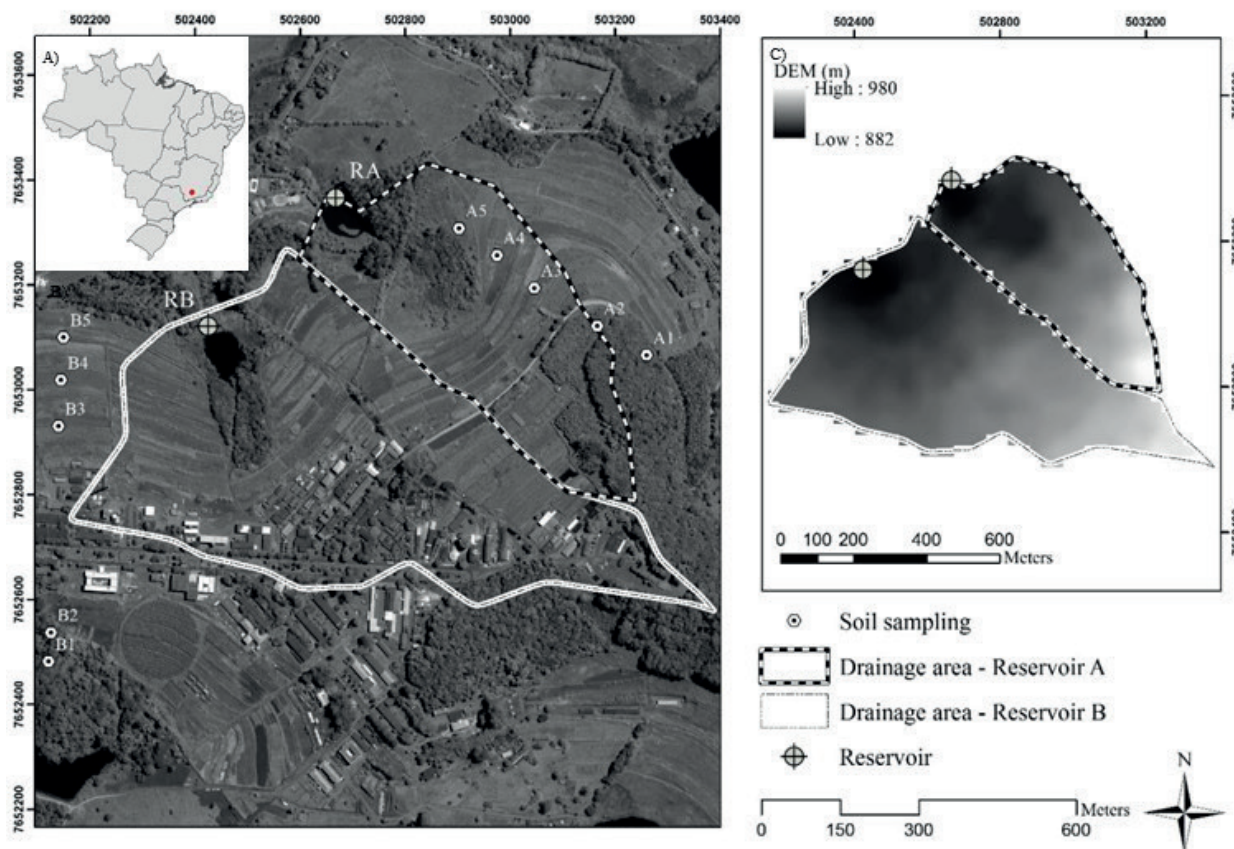
Therefore, this research aims to add to our knowledge of how magnetic sediment tracing performs in subtropical environments. In this work, we have, for the first time in Brazil, used magnetic properties as a tracer to characterize soil and to link upstream areas to reservoir sedimentation. Specifically, the objectives were to identify the origin of sediments at the Upper Grande River Basin, a tributary of the Paraná River, Southeastern Brazil, by sampling sediments from the bottom of small reservoirs and comparing magnetic characteristics to those of highly eroded potential soils in upstream areas.

## Materials and Methods

The study area is in the Upper Grande River Basin, one of the main tributaries of the Paraná River, at Lavras, Minas Gerais State, Brazil (21.13° S, 44.58° E) (Figure 1A). The area’s climate is classified as Cwb, a subtropical highland climate or monsoon-influenced temperate oceanic climate with a dry winter and a rainy summer, according to the Köppen classification system (Alvares et al., 2013). The average annual rainfall is 1,530 mm and the mean annual temperature 19.4 °C (Dantas, Carvalho, & Ferreira, 2007).

The drainage area for the two reservoirs (reservoirs RA and RB) was determined by processing a 30 m resolution Digital Elevation Model (DEM) obtained from shuttle radar topographic mission (SRTM) imagery. Flow direction and flow accumulation were calculated using the hydrology toolset for ArcGIS 10.1 (Environmental Systems Research

Institute [ESRI], 2010). Four points were assigned to the cell with the highest flow accumulation within the reservoirs and the watershed function from the hydrology toolset delimited the respective drainage areas (Figure 1B and 1C). The catchment areas for reservoirs A and B were 80,000 m<sup>2</sup> and 175,000 m<sup>2</sup>, respectively.



**Figure 1.** A) Drainage area location in the Brazilian territory, B) Soil sampling in the catenas, drainage area for the two reservoirs (RA and RB), and C) Digital Elevation Model (DEM) of the study area in Southeastern Brazil.

The two dominant soils in the catchment are Typic Hapludox and Anionic Acrudox (E. Silva, 2018). Both highly eroded potential (A. M. Silva et al., 2005; Lima et al., 2018) soils were sampled through representative catenas under different land uses: native forest, crop, eucalyptus and native pasture (Table 1). Due to issues with permissions, the soils from catena B were located just outside the watershed boundary. The soils were classified

according to the US Soil Taxonomy (Soil Survey Staff, 2014). At each soil sampling site, undisturbed triplicate samples up to 1 m depth, using a 50 mm diameter PVC (Polyvinyl chloride) core, were collected. Soil profile cores were sliced into 100 mm layers to enable measurement of the magnetic properties in each layer.

Sediment cores were extracted from near the embankment of both reservoirs using up to 50 cm

deep, 50 mm diameter PVC cores. The sediment cores from each reservoir were then sliced into 20 mm layers to enable the measurement of the magnetic properties in each layer, resulting in 15 samples for reservoir A and 18 for reservoir B.

Standardized soil color notation, using Munsell soil-color charts (Munsell Color Company, 1946), were used in order to determine colors of the sediments from reservoirs in order to ensure a non-iron oxides reducing environment.

**Table 1**  
**Catenas characterization**

Catena	Parent material	Land Use	Elevation (m)	Soil <sup>2</sup>
A	Gneiss <sup>1</sup>	Native Forest	940	A1
		Eucalyptus	933	A2
		Crop	922	A3
		Crop	912	A4
		Native Pasture	906	A5
B	Gabbro	Native Forest	909	B1
		Native Forest	919	B2
		Crop	909	B3
		Crop	903	B4
		Crop	893	B5

1. At the highest altitude, there is some contribution of gabbro.

2. A1: Rhodic Kandudult; A2, A3, A4 and A5: Typic Hapludox; B1, B2, B3, B4 and B5: Anionic Acrudox.

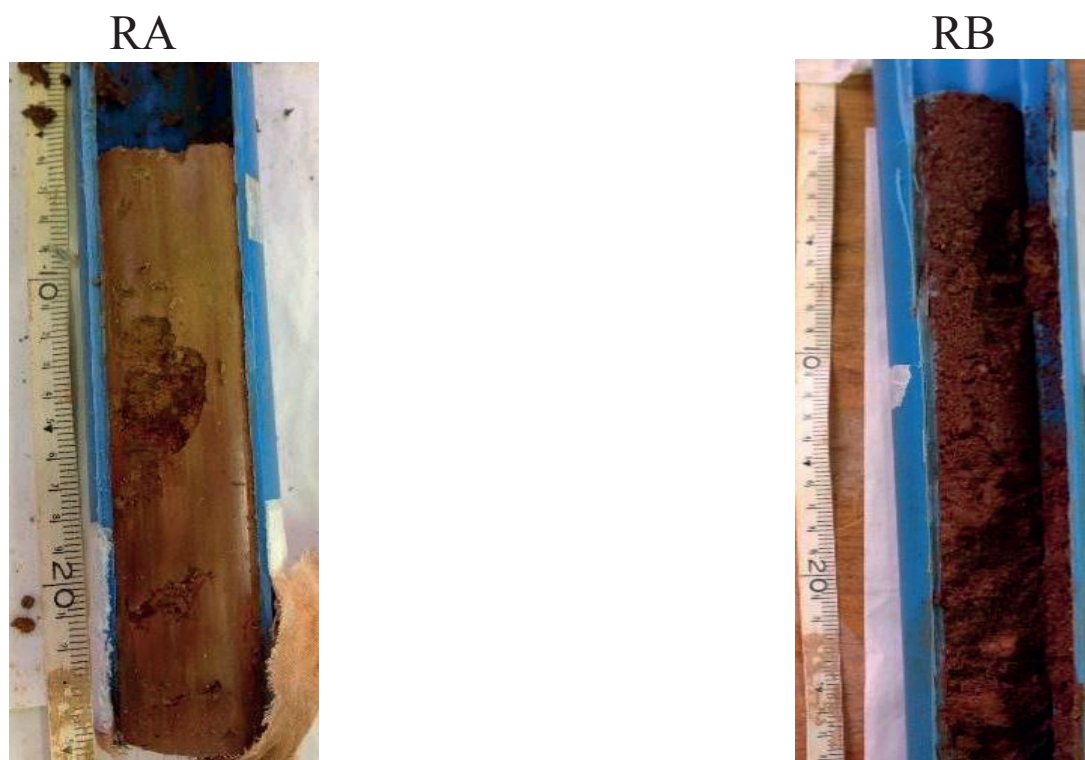
As magnetic properties are strongly particle-size dependent (Maher, Thompson, & Hounslow, 1999; Fontes, Oliveira, Costa, & Campos, 2000; Hatfield & Maher, 2008; Armstrong et al., 2012; Laceby et al., 2017), all of the samples (both soil and sediment) were separated into three particle size fractions (sand, silt and clay) prior to analysis. Samples were then treated with 1 mol L<sup>-1</sup> NaOH solution (10 g of oven-dried soil and 10 mL of sodium hydroxide (NaOH) 1N) and then moved to an ultrasonic bath in order to enhance soil particle dispersion (Claessen, 1997). The samples were wet sieved to obtain the sand size fraction (2-0.05mm). The remaining material was settled in Atterberg columns and separated into silt (0.05–0.002mm) and clay (<0.002mm) fractions. The separated fractions were dried at 40°C. For mineral identification of the clay, silt and sand fractions of soils, 0.3 g of each sample was analyzed by X-ray Diffraction (XRD) using the powder

method over the range of 5-50°2θ on a Bruker D2 Phaser diffractometer with Cu-Kα radiation, an Ni filter, a voltage of 30 kV and a current of 20 mA. For the magnetic analysis, 3g of each sample was packed into 10 cc plastic sample pots prior to magnetic analyses. It was not possible to measure the magnetic properties of the silt particle size in both soil and sediment due to the small amount of material. The magnetic parameters measured were low frequency magnetic susceptibility ( $\chi_{LF}$ ) and high frequency magnetic susceptibility ( $\chi_{HF}$ ). Magnetic susceptibility was measured at 0.47 kHz (low frequency) and at 4.7 kHz (high frequency) on a Bartington MS2B Susceptibility Sensor. From these measurements, frequency dependent susceptibility ( $\chi_{FD\%}$ ) was calculated (Dearing, Bird, Dann, & Benjamin, 1997). Measurements are expressed on a mass-normalized basis.

## Results and Discussion

Magnetic minerals can be very stable, which allows soil tracing. However, environmental conditions can modify the stability of these magnetic minerals (Snowball & Thompson, 1988; Walden et al., 1999). To rule this out, standardized soil colour notation, using Munsell soil-colour charts (Munsell Color Company, 1946), were compared to the sediment color. Sediment hues were similar to the

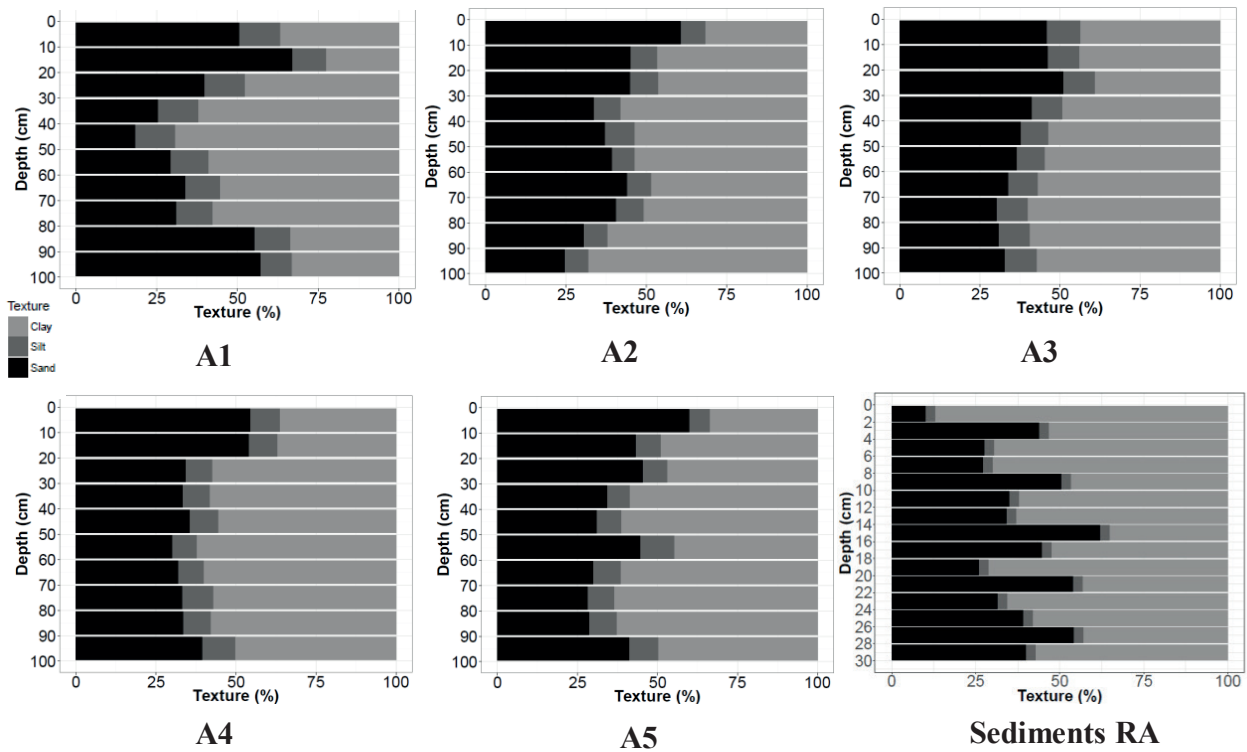
original soil profiles: 10YR (red-yellow) in soils and sediments of RA, and 2.5YR (red) in soils and sediments of RB. These trends are consistent with oxidative (well-aerated) environments (Resende, Curi, & Rezende, 2017) attributable to a constant water flow through the reservoir (Figure 2). Consequently, changes to the magnetic properties of the sediment due to oxide reduction is unlikely.



**Figure 2.** Sediment profile cores from reservoirs A (RA) and B (RB).

The particle size distribution for each sample site is illustrated in figures 3 and 4. The Catena A soils present less clay content (50%) than Catena B soils (up to 80% of clay). This is explained by the much smaller amount or absence of quartz (a very resistant

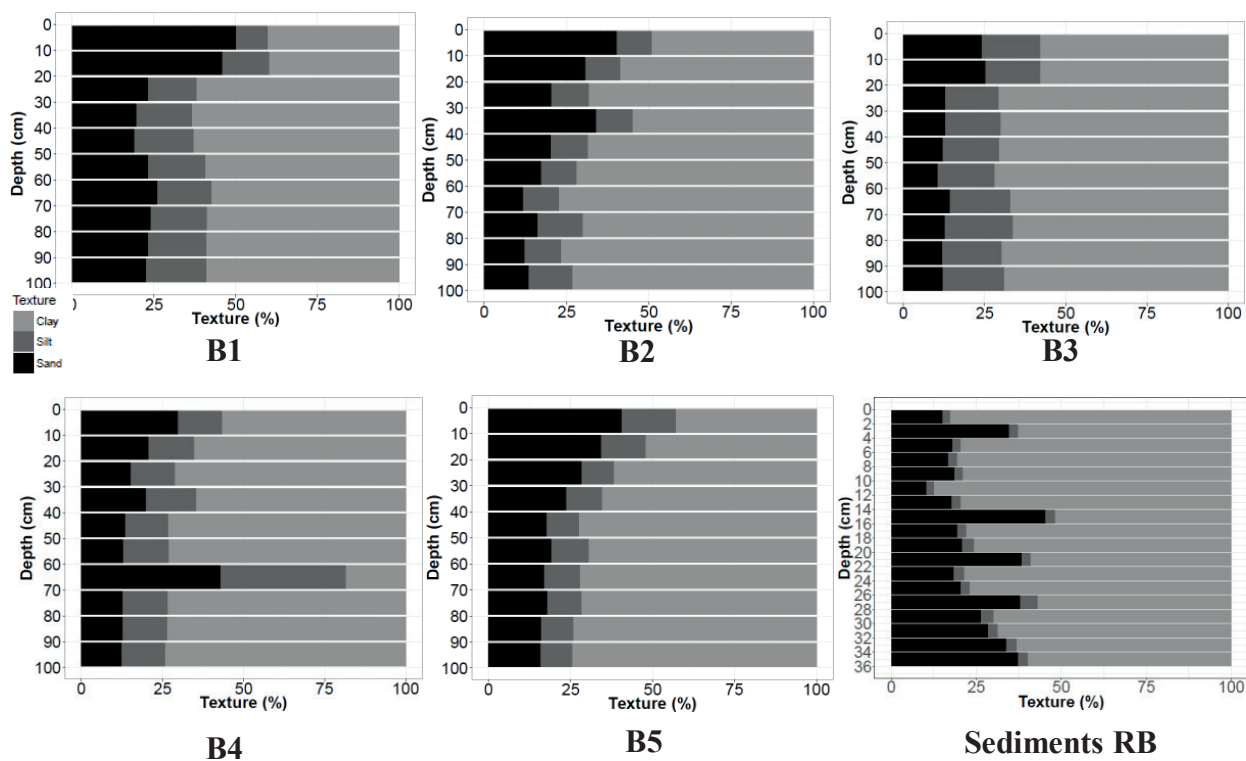
primary mineral) content in gabbro compared to gneiss (Ker, 1997; Curi & Kämpf, 2012; Kämpf, Marques, & Curi, 2012), and supported by the soils' x-ray diffraction analysis (Figures 5, 6 and 7).



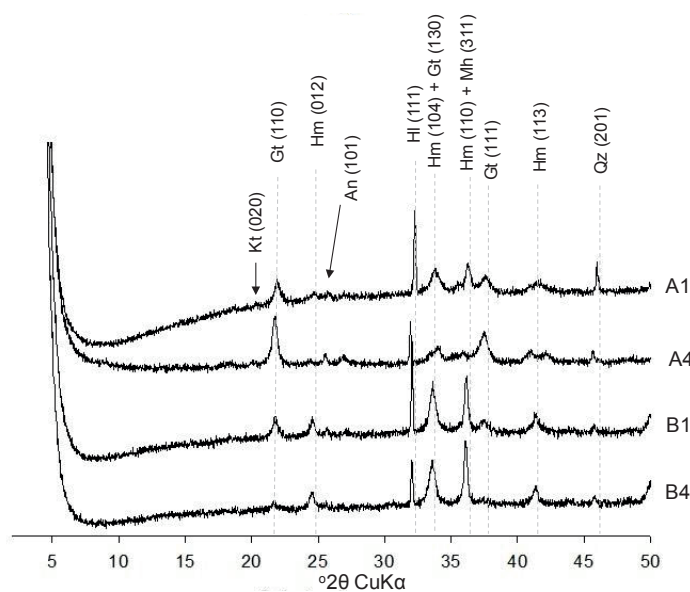
**Figure 3.** Particle size distribution of Catena A soils, to 1 meter depth, divided in 10 cm layers, at each site, and of reservoir RA, to 0.3 meter depth, divided in 2 cm layers. Soils A1, A2, A3, A4 and A5 characterizations are presented at Table 1.

In addition to differences in particle size distributions, there were also significant differences in the magnetic constituents of the different catenas and size fractions. Much higher ferrimagnetic maghemite content was detected in the clay fraction of the Catena B soils than the Catena A soils,

reflecting the gabbro influence in soil B and its lack of influence in soil A (Figure 5). Such different mineral content in the soils sampled resulted in differences in magnetic measurements (Figures 7, 9, 10, 11 and 12).



**Figure 4.** Particle size distribution of Catena B soils, to 1 meter depth, divided in 10 cm layers, at different sites, and of reservoir RB, to 0.3 meter depth, divided in 2 cm layers. Soils B1, B2, B3, B4 and B5 characterizations are presented at Table 1.

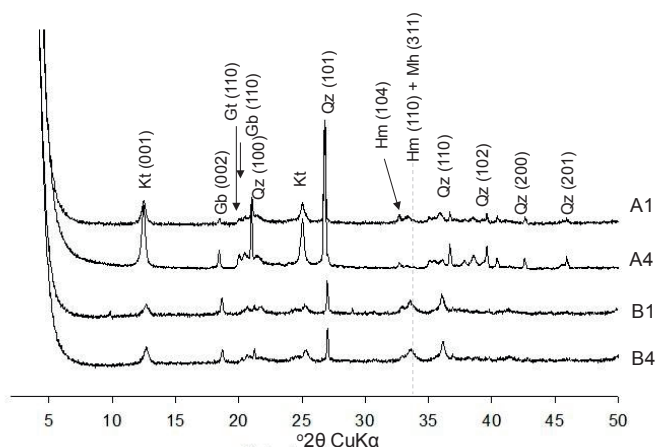


**Figure 5.** X-ray diffraction of powder samples of the Fe-concentrated clay fraction of soils. 10% of halite was added as an internal standard. Kt: kaolinite; Gt: goethite; Hm: hematite; An: anatase; Hl: halite; Mh: maghemite; Qz: quartz. Soils A1, A4, B1 and B4 characterizations are presented at Table 1.

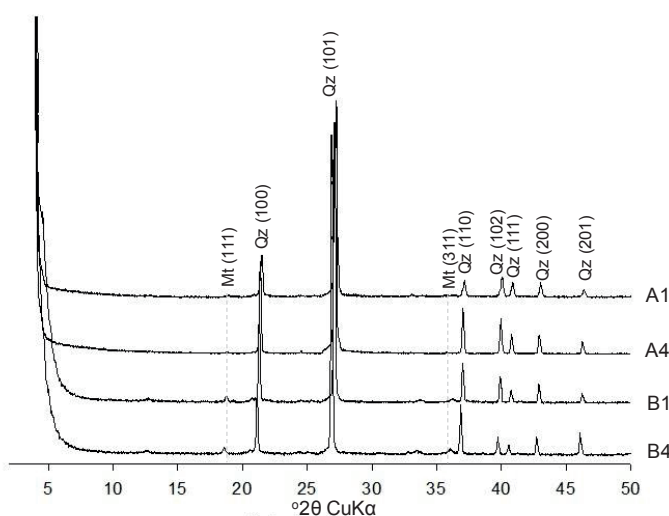


Maghemite was identified in the silt fraction of Catena B soils (Figure 6) and magnetite was found in the sand fraction (Figure 7). Consequently, values of low frequency magnetic susceptibility ranging from 58.99 to 601 x 10<sup>-7</sup> m<sup>3</sup> kg<sup>-1</sup> were measured in the Catena B soil samples (Figures 8 and 9); soils developed from mafic rocks are often characterized by higher  $\chi_{LF}$  values (Costa, Bigham, Rhoton, & Traina, 1999; Dearing, 1999; Lu, Xue, Zhu, & Yu, 2008; A. R. D. Silva, Souza, & Costa, 2010).

Apart from the small content of maghemite in the clay fraction of soil A1, no other magnetic minerals were identified in the Catena A soils, which is due to the gneiss parent material. Consequently, the Catena A soils, which formed over gneiss, are largely dominated by diamagnetic minerals (quartz, feldspars and muscovite) which are characterized as having no or very low magnetic susceptibility (Walden et al., 1999);  $\chi_{LF}$  values ranged from 2.43 to 31.91 x 10<sup>-7</sup> m<sup>3</sup> kg<sup>-1</sup> (Figures 8 and 9).



**Figure 6.** X-ray diffraction of powder samples of the silt fraction of soils. Kt: kaolinite; Gb: gibbsite; Gt: goethite; Qz: quartz; Hm = hematite; Mh: maghemite. Soils A1, A4, B1 and B4 characterizations are presented at Table 1.

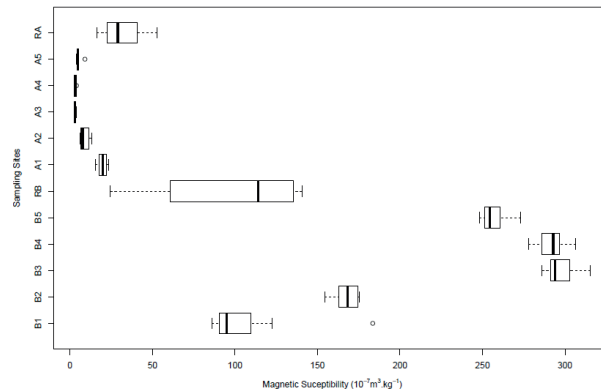


**Figure 7.** X-ray diffraction of powder samples of the sand fraction of soils. Mt: magnetite; Qz: quartz. Soils A1, A4, B1 and B4 characterizations are presented at Table 1.

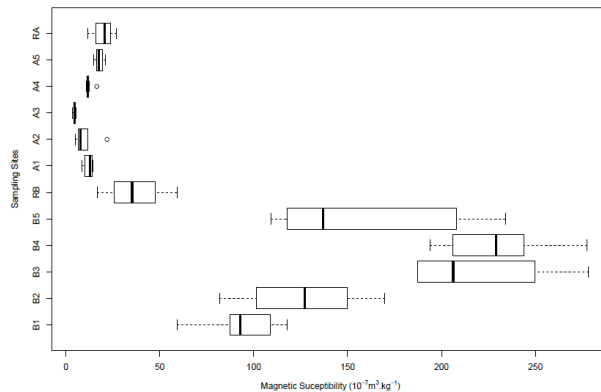
The difference in low frequency magnetic susceptibility between Catena B samples and Catena A samples, due to the greater content of magnetic minerals in Catena B samples, makes low frequency magnetic susceptibility an effective basis for source differentiation, as found in other studies (Pulley & Rowntree, 2016). Low frequency magnetic susceptibility is mainly associated with the Fe oxides assemblage and content within a soil: magnetite ( $\text{Fe}_3\text{O}_4$ )  $\chi_{\text{LF}}$  values can be found in literature ranging from 4,000 to 10,000  $10^{-7} \text{ m}^3 \text{ kg}^{-1}$ , while maghemite ( $\gamma\text{Fe}_2\text{O}_3$ ) varies from 2,500 to 4,500  $10^{-7} \text{ m}^3 \text{ kg}^{-1}$  and goethite ( $\alpha\text{FeOOH}$ ) varies from 3 to 13  $10^{-7} \text{ m}^3 \text{ kg}^{-1}$  (Walden et al., 1999). The magnetic susceptibility results obtained here are comparable to those reported by S. Silva et al. (2016) when evaluating the efficiency of a magnetometer

as a tool for mapping soil classes and properties in tropical conditions, with  $\chi_{\text{LF}}$  values for Hapludox up to  $260 \times 10^{-7} \text{ m}^3 \text{ kg}^{-1}$ . The  $\chi_{\text{LF}}$  values measured in the Catena A soil samples are lower than those observed by Poggere et al. (2018), ranging from 73 to 307  $10^{-7} \text{ m}^3 \text{ kg}^{-1}$  due to lower Fe oxides content in the present research (Figures 5, 6 and 7).

Figures 8 and 9 summarize some of the observed magnetic contrasts, making sediment source identification possible by comparing both soils' sediment signatures with those of the reservoirs' sediments. Systematic variation in the horizons down the profile were observed, with reservoir B presenting values that indicate influence from sediments of both catenas due to the mixture of different components (Figures 5, 6 and 7).



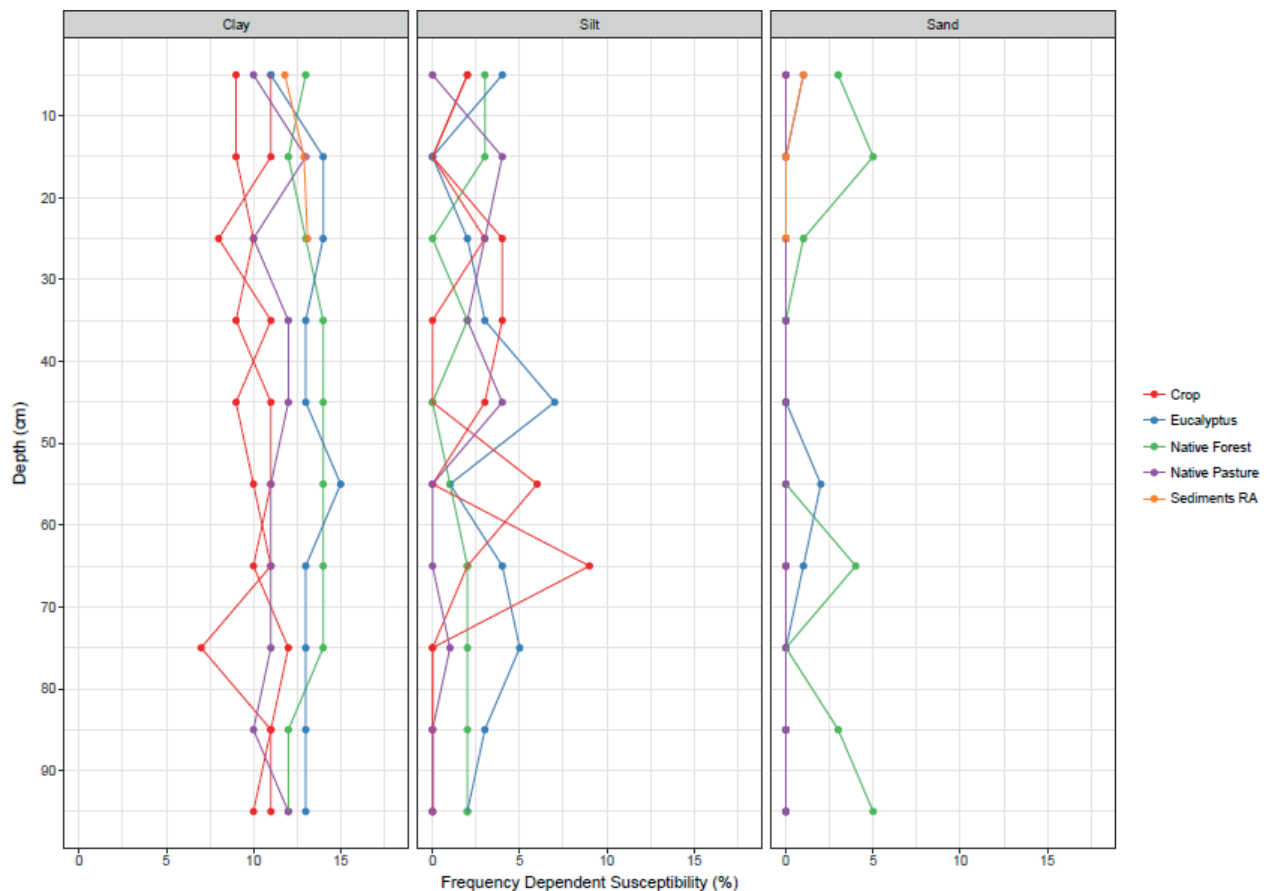
**Figure 8.** Low frequency magnetic susceptibility ( $10^{-7} \text{ m}^3 \text{ kg}^{-1}$ ) of clay fraction of soil profiles, with different land uses (Table 1).



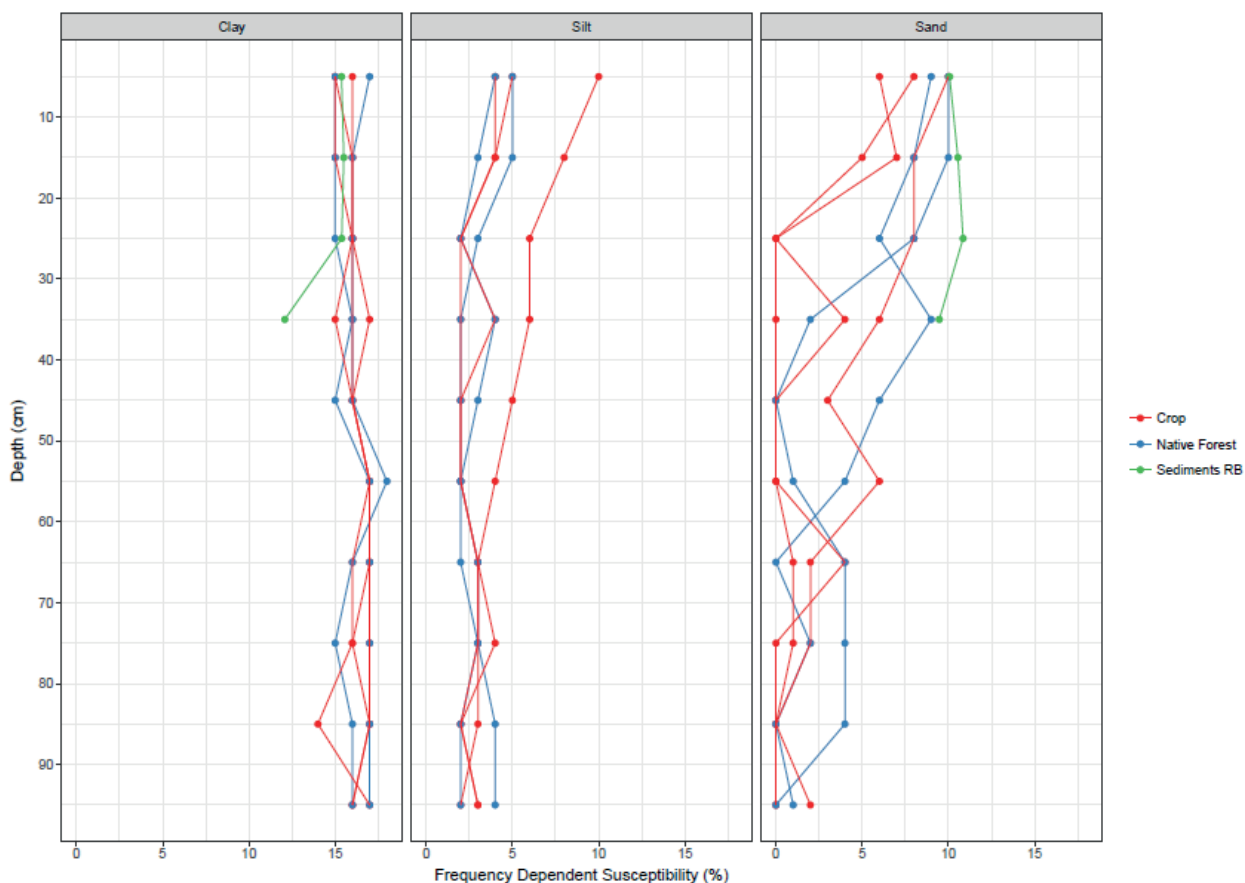
**Figure 9.** Low frequency magnetic susceptibility ( $10^{-7} \text{ m}^3 \text{ kg}^{-1}$ ) of sand fraction of soil profiles, with different land uses (Table 1).

Figures 10 and 11 illustrate the strong particle size dependence of magnetic behavior of the soils/sediments. As Catena B soils present a high  $\chi_{FD\%}$  (10-14%) in the clay and silt fractions, superparamagnetic and stable single domain grains are likely. Such magnetic behaviors are evidenced by the in situ formation of maghemite, most likely by oxidation of lithogenic magnetite during soil formation (Curi, 1983; Costa et al., 1999; Dearing,

1999; Walden et al., 1999; Hatfield & Maher, 2009; Pulley & Rowntree, 2016). In the sand fraction of Catena B soils, magnetite presence justifies their higher  $\chi_{FD\%}$  in comparison with Catena A soils, which present a virtual absence of superparamagnetic grains (Dearing, 1999; Kämpf & Curi, 2000). The data variation emphasizes the need for magnetic characterization of sediments and possible sources on a particle size basis (Hatfield & Maher, 2008).



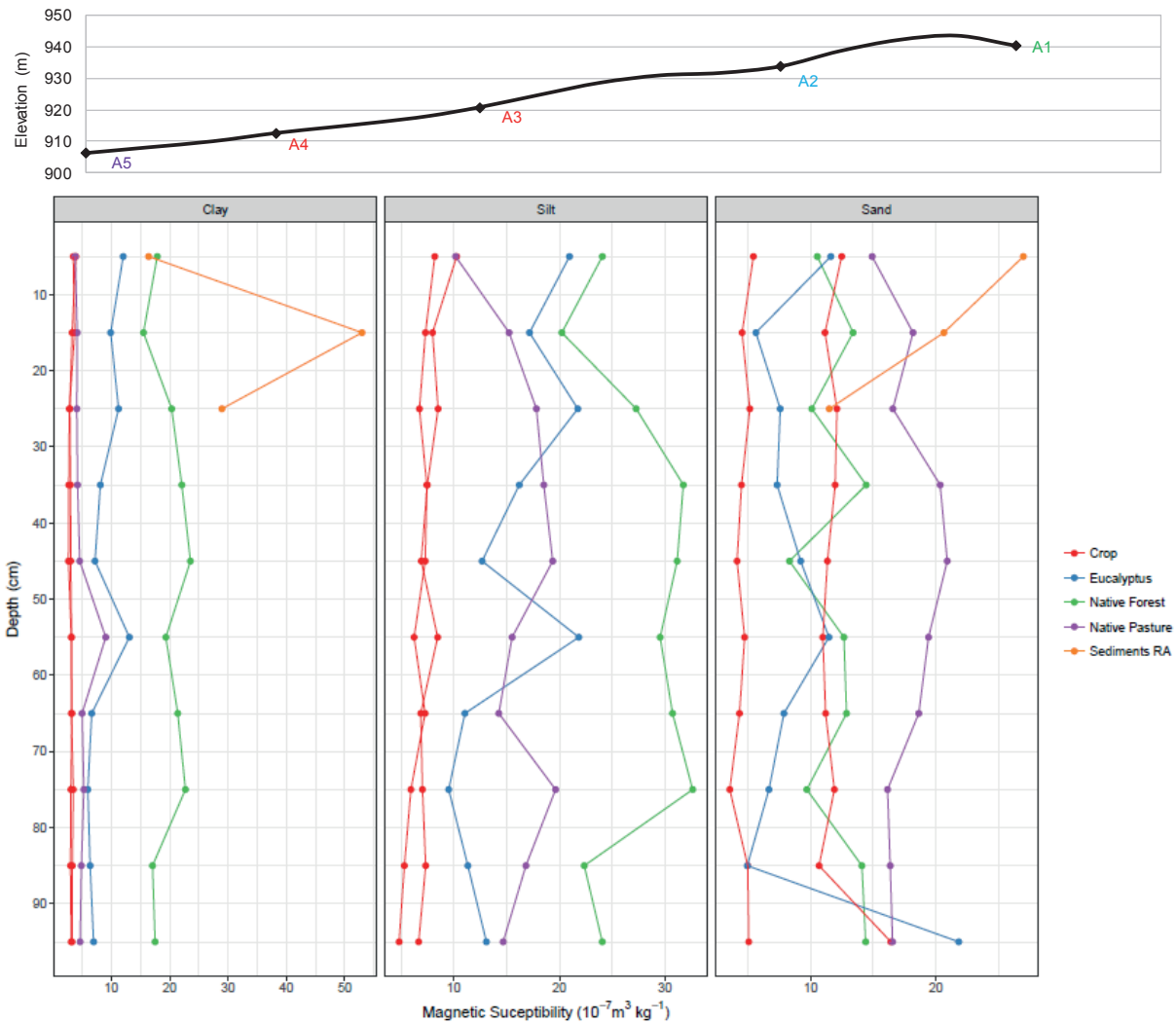
**Figure 10.** Frequency dependent susceptibility (%) of the different particle sizes of Catena A soil profiles, with different land uses, and in sediments of reservoir A.



**Figure 11.** Frequency dependent susceptibility (%) of three different particle sizes of Catena B soil profiles, with different land uses, and in sediments of reservoir B.

The elevation (in meters) of the soils sampled under different land uses, as well as  $\chi_{LF}$  ( $10^{-7} \text{ m}^3 \text{ kg}^{-1}$ ) of different particle sizes at 1 meter down soil profiles, are presented in figures 12 and 13. The magnetic measurements appear to differentiate the two major potential suspended sediment inputs (Catena A and B) to both reservoirs (RA and RB): figure 12 indicates a dominant Catena A sediment source at reservoir A, corroborating with the relief

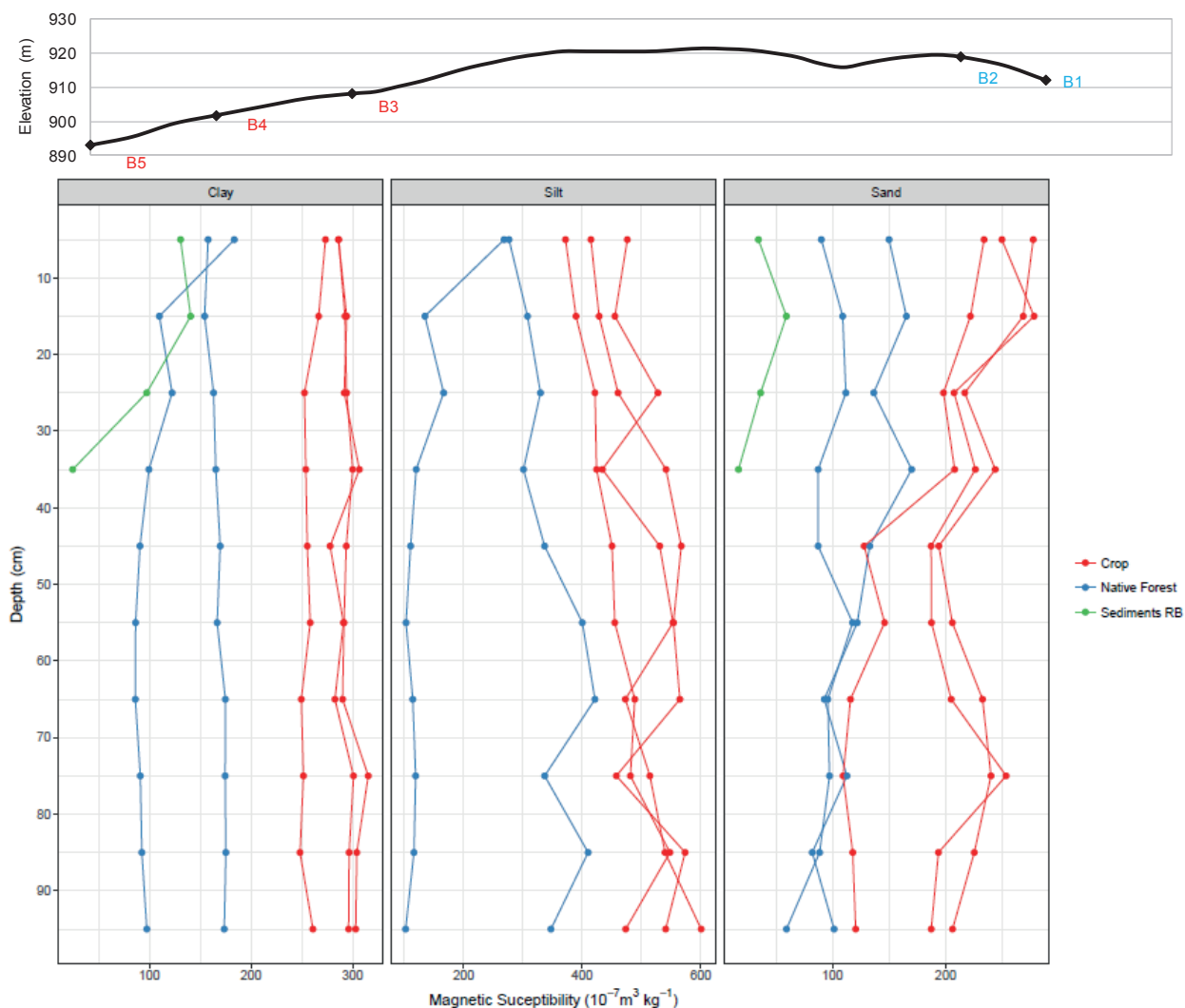
and consequent flow direction of the drainage area indicated in figure 1; on the other hand, Catena B soil samples look like the main source of sediment contribution at reservoir B. Soils with higher clay content, as Catena B soils, tend to produce higher sediment yield given the preferential transport of finer (smaller diameter) and lighter (lower density) sediment by the water erosion process (Morgan, 2009; Vahabi & Nikkami, 2008).



**Figure 12.** Elevation (m) of sampling sites and low frequency magnetic susceptibility ( $10^{-7} \text{ m}^3 \text{ kg}^{-1}$ ) (note differing scales) of different particle sizes of Catena A soil profiles, with different land uses, and in sediments of reservoir A.

The similarity of Catena A soil clay magnetic susceptibility values to those of the reservoir sediments (Figures 12 and 13) can be indicative of a higher erosion rate in Catena A than Catena B soils, as stated by Lima et al. (2018) and A. M. Silva et al. (2005), respectively. This finding is explained by the higher water infiltration in Catena B soils, a consequence of the granular structure compared with the blocky structure of Catena A soils, which promotes lower water infiltration and consequently higher runoff (Ferreira, Fernandes, & Curi, 1999). Also, the Catena A slope was steeper and crusting was observed during soil sample collection,

both of which increase the potential for erosion. Further, within Catena A soils, higher erosion rates of clay were indicated for A1 soils (Figure 12), an Ultisol, whose A horizon presents weaker structure development and lesser clay content than B horizon. Within Catena B soils, higher erosion rates were associated with B1 and B2 soils (Figure 13), which occur on steeper slopes than other soils of this catena. As soils, sediments and rocks from different locations in a catchment are characterized by different magnetic properties (Cervi, Maher, Poliseli, Souza, & Costa, 2019), and so such a technique can be used on different soil classes.



**Figure 13.** Elevation (m) of sampling sites and low frequency magnetic susceptibility ( $10^{-7} \text{ m}^3 \text{ kg}^{-1}$ ) (note different scales) of different particle sizes of Catena B soil profiles, with different land uses, and in sediments of reservoir B.

The analysis of different soil properties (standardized soil color notation, particle size distribution, x-ray diffraction analysis, low frequency magnetic susceptibility and frequency dependent susceptibility) enhance magnetic contrasts that make sediment source identification possible. Soil properties combination analysis has been widely used as a fingerprint method to assess sediment sources (Pulley, Van der Waal, Rowntree, & Collins, 2018; Tian et al., 2019). The qualitative assessment allows sediment source identification and enables a future sustainable soil and water conservation management.

## Conclusions

This research provides evidence that sediment tracing using magnetic parameters to identify possible sediment sources in a subtropical environment, specifically in Brazil, is a useful technique to delineate soil erosion and sediment transfer.

The strong difference in magnetic signatures is a primary result of different parent materials, since the climate and major pedogenic process are quite similar (desilication and residual concentration of oxides). The different parent materials lead to

soils with different contents and types of Fe oxide minerals, inducing differential magnetic variability, which allows the identification of sources of deposited sediments. Different behavior of soil types could be clearly distinguished by magnetic property evaluation and the origin of sediments of downstream reservoirs allocated to their source. Given the low cost of magnetic measurements, and distinct properties among soil classes, magnetic tracing offers significant potential in soil erosion and sediment tracing studies in the tropics. This increased understanding could be used to develop appropriate soil erosion mitigation methods, maintaining soil quantity and quality and avoiding detrimental impacts in the downstream aquatic systems.

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